

# Grizzly bear population genomics across a coastal-interior ecotone in British Columbia, Canada

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# Abstract

Local adaptation research often focuses on discrete populations without extensive gene flow that are under differential selective pressures. By contrast, grizzly bears *Ursus arctos* in British Columbia (BC) are wide-ranging omnivores that span an environmental and resource ecotone from the coastal, salmon-enriched rainforest to dry interior plateau. This ecotone has been associated with local adaptation in other species and the different regions to morphological variation in grizzly bears. To understand genome-wide population genetic patterns across the ecotone and to identify loci or genomic regions associated with these different environments, here we use whole-genome resequencing to characterize 3.9 M SNPs in 31 grizzly bears spanning central to northern latitudes in coastal and interior regions (to the west and east of the Coastal Mountain Range (CMR), respectively). Clustering grizzly samples by genotypes identified three groups that generally correspond to the source geographic regions, with the greatest variation occurring from north to south. The data were best explained by a single ancestry cluster, but  $K = 3$  recovered the three geographic groupings and was used to identify putative non-migrant individuals. The presence of individuals with mixed ancestry (using  $K = 3$ ) provides evidence for travel across the CMR, but significant differentiation between clusters (mean  $F_{ST} = 0.015-0.036$ ) suggests some genetic separation between the regions, supporting an isolation-by-distance or clinal variation model. Putative close-kin were identified and removed, then multiple supervised outlier SNP detection methods were applied to identify regions of the genome consistently segregating between coastal and interior regions. Several associated genomic regions and candidate genes were identified, including a consistently identified outlier region near the gene *creatine kinase, m-type*. This work provides the first genome-wide analysis of grizzly bears in the studied region. These findings will be useful for connectivity planning and research on the adaptability of coastal and interior grizzlies to future climate change scenarios.

**Keywords:** conservation genomics; grizzly bear; local adaptation; population genomics; single nucleotide polymorphism; whole genome resequencing

## Introduction

Local adaptation can occur due to evolutionary processes that provide fitness advantages to different populations in response to local environmental pressures (Kawecki and Ebert 2004; Blanquart et al. 2013), for example as a result of elevation (Halbritter et al. 2018) or salinity gradients (Sanford and Kelly 2011). This can be impacted by gene flow, as demonstrated in host-parasite interactions, where the relative rate of gene flow in host and parasite can determine local adaptation outcomes (Hoeksema and Forde 2008). Although local adaptation is typically documented in case studies with limited or no gene flow, it can also occur at microgeographic scales and with gene flow among populations (Richardson et al. 2014; Tigano and Friesen 2016). In some cases, specific regions of the genome can be associated with significant phenotypic differences even though the rest of the genome is undifferentiated (e.g., run timing; Barry et al. 2024), and this highlights the importance of considering genome-wide data when investigating local adaptation.

In the presence of gene flow, ecotones (i.e., transition areas between ecological communities) can foster local adaptation (Wright et al. 2006; Yuan et al. 2018; Ekar et al. 2019). For example, substrate coloration differences along the ecotone of White Sands, New Mexico (USA), where the white gypsum geological feature transitions from lighter to darker soil, resulted in dorsal color variation within populations of three local lizard species (Rosenblum 2006). In reciprocal transplantation experiments of eastern oyster (*Crassostrea virginica*) along the lagoon ecotone in eastern Florida (USA), reciprocal home-site advantage occurs and signatures of local adaptation are present (Burford et al. 2014). Ecotone transition zones can have pronounced environmental gradients, and this can exert different selective pressures depending on the location across the gradient in which the population resides.

Detecting local adaptation has become increasingly possible with rapid advances in genomics (Hoban et al. 2016). Although local adaptation was previously investigated by reciprocal transplants (an approach not possible for all species), genomic datasets now allow the detection of genomic patterns underlying adaptive diversity (Funk et al. 2019). Loci that are putatively adaptive can be identified through association tests of allele frequencies and environmental variation. When environmental drivers of local adaptation are unknown, genome scans can contrast across populations to identify candidate outlier loci with elevated

differentiation (Hoban et al. 2016). This is possible with relatively few samples, but understanding the role of the associated loci often requires an investigation in a larger proportion of the population (Lotterhos and Whitlock 2015). Genome scans have identified signatures of local adaptation in response to habitat differentiation, latitudinal gradients, and environmental clines in many species, including the Eurasian blue tit *Cyanistes caeruleus* (Perrier et al. 2020), the Mediterranean striped mullet *Mullus surmuletus* (Dalongeville et al. 2018), and the thick-billed murre *Uria lomvia* (Tigano et al. 2017).

A dramatic ecotone in British Columbia (BC), Canada, spans the coastal rainforest environment through the dry interior plateau and provides the necessary conditions to foster intraspecific variation. These disparate environments are divided by the Coast Mountain Range (CMR). The CMR hosts a hybrid zone for subspecies of the Swainson's thrush *Catharus ustulatus* (Ruegg 2008), a region of intraspecific migratory, morphological, and genetic variation in the Hermit thrush *Catharus guttatus* (Alvarado et al. 2014), and an area of introgression between white spruce *Picea glauca* and Sitka spruce *P. sitchensis* (Bennuah et al. 2004). Even highly mobile mammals show divergence across this ecotone; the grey wolves *Canis lupus* of coastal BC are highly differentiated from interior populations and represent a unique ecotype and evolutionarily significant unit (Muñoz-Fuentes et al. 2009; Schweizer et al. 2016b) with genetic differences in genomic regions related to dietary fat and lipid metabolism and coat colour (Schweizer et al. 2016a). Other differences exist in grey wolves across this ecotone, including divergent mitochondrial genetics and differences in morphology and diet (Muñoz-Fuentes et al. 2009), where the coastal wolves forage extensively on salmon and other marine resources (Darimont et al. 2003). A major segregating environmental resource for the coastal-interior ecotone is from the various dietary resources available to generalist species (e.g., abundance and availability of salmonids *Salmonidae* spp.).

Salmonids are a defining resource for many species, ecosystems, and people in the coastal region of BC. Grizzly bears *Ursus arctos* are particularly reliant on salmon. Individual bears with increased access to salmon tend to be larger, have bigger litters, have lower stress hormone levels, and exist at higher densities at a population level (Rausch 1963; Hilderbrand et al. 1999; Bryan et al. 2013). Salmon also migrate into the interior (Quinn 2018), but contemporary salmon availability and consumption declines significantly past the CMR (Hilderbrand et al. 1996; Adams et al. 2017; Adams et al. 2024). Phenotypic differences occur

across the coastal-interior ecotone in grizzlies, with interior grizzlies having smaller body mass and skull size than coastal individuals (Rausch 1963; Kurtén 1973; Paetkau et al. 1998). The potential causes of this phenotypic difference are not fully understood. It may be related to phenotypic plasticity and salmon consumption (Hilderbrand et al. 1999); larger males require high-protein food to gain mass (Robbins et al. 2007) and smaller bears can gain mass through vegetation consumption alone (Felicetti et al. 2003). Alternately, since interior bears have long existed without widely-available aggregations of high-quality food like salmon, they may have adapted to intermittent availability of meat resources contrasting the coastal grizzlies that depend on salmon to support their larger body mass (Mowat and Heard 2006). Size differences between some coastal and interior populations could also be influenced by differential introgression of polar bear *U. maritimus* alleles through past hybridization events (Cahill et al. 2015; Cahill et al. 2018; Miller et al. 2024), although a potential role for this in coastal BC has not been described. The clinal variation in resources and associated morphology in grizzly bears across the BC coastal-to-interior ecotone makes it a valuable system to investigate genomic signatures of local adaptation in this wide-ranging species.

Population genomics and local adaptation can be used to inform management and policy (Waples et al. 2022), and is expected to support the assessment of the provincially managed Grizzly Bear Population Units (GBPUs; Province of British Columbia 2012), and management activities of Indigenous Stewardship Offices. Local Indigenous ecological knowledge indicates that interior bears migrate from the interior to the coastal areas through the Bella Coola Valley, Nuxalk Territory (e.g., Stuie; 52.3699°N, 126.0659°W) to access resources including salmon (Jason Moody, Nuxalk Nation, *pers. comm*). Therefore, we hypothesize that the Bella Coola Valley and other valleys like it that transect the CMR may provide opportunities for gene flow across the ecotone. Furthermore, the ecotone may provide environmental resource gradients that could foster local adaptation at the extremes of each region, and so here we aim to identify genomic loci or regions of elevated segregating genetic variation between the coastal and interior regions. Collectively, this work provides new genomic resources and insights regarding the population genetics of grizzly bears in BC, and on the presence of and genomics underlying putative local adaptation across the coastal-to-interior ecotone. This work also provides insights for future work regarding the relevance of adaptive variation in grizzly bear conservation and resilience to future climate scenarios.

## Methods

### *Sample collection, DNA extraction, library preparation and sequencing*

Dried hide samples were obtained from the BC Ministry of Environment compulsory inspection database of killed grizzly bears through a data share agreement. Samples were selected from the Central Coast and adjacent interior region to represent the coastal-to-interior plateau ecotone of BC (Figure 1). Samples were collected from 1996 to 2016. Geographic coordinates provided per sample are slightly offset (jittered) locations of where each bear was killed (Additional File S1), as per requirements of the data share agreement. The recorded phenotypic sex of the selected samples ( $n = 31$ ) included nine females and 22 males, a sex ratio that reflects the higher frequency of males in killed bears.

Genomic DNA was extracted from hide tissue using the DNeasy Blood and Tissue kit (QIAGEN) from 25 mg slivers of dried tissue following manufacturer's guidelines, but with an overnight incubation in lysis buffer and a second separate elution from the columns. Purified genomic DNA was quantified using spectrophotometry (Nanodrop; ThermoFisher) and fluorimetry by Qubit dsDNA-BR (ThermoFisher). Samples were submitted to the Génome Québec Innovation Centre for PCR-free shotgun whole-genome library preparation to be sequenced on HiSeqX and NovaSeq6000 S4 (Illumina) using paired-end 150 bp reads to a per-sample target depth of 10X coverage.

### *Variant calling and filtering*

Variants were called using the GATK pipeline (Van der Auwera et al. 2013) as described below and in the associated code repository (see Data Availability). Paired-end reads were aligned to the *Ursus arctos* reference genome (GCF\_003584765.1; Taylor et al. 2018) using *bwa mem* (v.0.7.17; Li 2013). Read groups were added with experimental information using the Picard Toolkit function *AddOrReplaceReadGroups* (v.2.18/9; Broad Institute 2024). Alignments were sorted and indexed using SAMtools (v.1.9; Danecek et al. 2021). Alignment rates were calculated based on the number of alignments passing a minimum threshold of at least 100 bp alignment with a minimum percent identity of 98%, expressed as a fraction of the total aligned reads (see Data Availability). PCR duplicates were identified and marked using the Picard

Toolkit function *MarkDuplicates*, and samples that were split between sequencing lanes were merged based on read group identifiers.

All GATK batch scripts are provided (see Data Availability) and described here in brief. Haplotypes were called with the GATK *HaplotypeCaller* using flags *genotyping\_mode* DISCOVERY and *emitRefConfidence* GVCF. Genotypes were extracted from the resulting GVCF files using the function *GenotypeGVCF* at intervals of 10 Mbp. The resultant files were merged using the function *CatVariants*. All merged files were sorted using the *vcf-sort* function of VCFtools (Danecek et al. 2011). Variants were then scored using the *VariantRecalibrator* in SNP mode, and *ApplyRecalibration* functions of GATK. Filtering of variants was conducted using VCFtools to remove indels and to only retain biallelic SNP variants with a minor allele frequency (MAF)  $\geq 0.05$ . SNPs were filtered to keep those with less than 10% missing data across samples.

After SNP calling was completed and during project analysis, a new chromosome-level reference genome became available (GCF\_023065955.2; UrsArc2.0; Armstrong et al. 2022). To make use of the updated genome, with improved metrics, annotation, and assembly of sex chromosomes, the program SNPlift (Normandeau et al. 2023) was used to transfer SNP positions from the reference genome used for calling SNPs to the new assembly (UrsArc2.0). The output VCF file was provided a new header with the bcftools function *reheader*. Variants oriented to the UrsArc2.0 assembly were used for all downstream applications.

SNPs transferred to UrsArc2.0 then underwent additional filtering to remove SNPs within 5 bp of indels, only keep SNPs with an overall quality score (i.e., QUAL) of at least 20 and an average read depth across all samples  $\geq 7$ . Subsequently, all genotypes per individual supported by  $< 5$  reads or  $> 1,000$  reads were set to missing values, as were any genotypes with individual quality scores (GQ)  $< 20$ ; SNPs were then filtered again to only retain SNPs missing in  $< 15\%$  of individuals. A final all SNPs dataset was generated by reapplying the MAF filter (MAF  $> 0.05$ ), and an LD-filtered dataset was generated for population genetic purposes by removing SNPs based on linkage (i.e., keep if linkage  $< 0.5$  in 50 kbp windows) using bcftools (Danecek et al. 2021).

# *Mitochondrial DNA haplotypes and phylogenetics*

To fit the samples from the current study into a broader phylogenetic context with previous work, the multiple alignment program MAFFT (v.7; Katoh et al. 2017) was used to align a 701bp fragment of the mitochondrial control region for each of the four unique haplotypes from the 31 samples, five haplotypes from coastal Alaska (Talbot et al. 2014), and all 80 previously published haplotype sequences representing grizzly bear mitochondrial clades 1- 6 from Miller et al. (2006) to the reference genome (Taylor et al. 2018). A phylogenetic tree was generated using the MEGAX program (Kumar et al. 2018) with the Maximum Likelihood method and the Tamura-Nei model (Tamura and Nei 1993) applying Neighbor-Join and BioNJ algorithms. An American black bear *Ursus americanus* control region sequence was used as the outgroup (Miller et al. 2006).

## *Genetic characterization, relatedness, and genetic sex*

The LD-filtered VCF file was used for population genetic characterization (see Data Availability). To avoid impacts of sex-linked loci (Benestan et al. 2017), SNPs on the sex chromosomes were removed using bcftools to create an autosome-only dataset (any SNPs on the mitochondrial genome were also removed). The VCF file was loaded into the R environment (R Core Team 2025) and converted to genind format using vcfr (Knaus and Grünwald 2017) to be analyzed using the *simple\_pop\_stats* repository (see Data Availability). The genind file was converted to genlight format to conduct a principal component analysis (PCA) through the *gi2gl* function of dartR (Gruber et al. 2018) followed by the *glPca* function of adegenet (Jombart and Ahmed 2011). PCA results were plotted using ggplot2 (Wickham 2016), with various sample metadata overlayed to inspect general trends, including percentage of missing data, genetic sex, geographic location of sampling, and year of sampling.

The dataset was converted to demerelate format with dartR, then converted to related format (Pew et al. 2015) using the related function *readgenotypedata*. The relatedness between individuals was calculated using the coancestry function of related, and the Ritland relatedness statistic (Ritland 1996) was used to interpret inter-individual relatedness. Outlier relatedness levels were determined by using *boxplot.stats* of R, and a relatedness cutoff value of 0.15 was used to consider pairs as having elevated relatedness. A single individual per pair with elevated



relatedness was removed from the dataset until no pairs above the cutoff remained. A PCA was generated as described above using the close kin-removed dataset.

The genetic sex of individuals was confirmed against phenotypic (recorded) sex by taking raw reads for all 31 samples and aligning against the UrsArc2.0 assembly (GCF\_023065955.2) using *bwa mem*, then sorting and indexing with SAMtools. The lengths of all chromosomes were calculated using custom python code (see Data Availability; E. Normandeau, Scripts), and the coverage across each chromosome was calculated using the `genomeCoverageBed` function of `bedtools` (Quinlan and Hall 2010). The average depths of coverage for the X and Y chromosomes were calculated, and the ratio of the coverage was determined to identify XX or XY individuals.

### *Population substructure and cluster membership*

The optimal number of clusters in the data were determined using several approaches. First, an unsupervised discriminant analysis of principal components (DAPC) was conducted, and the Bayesian information criterion (BIC) was observed for different numbers of clusters identified. A BIC elbow was sought, and scatterplots and assignplots were generated for two or three clusters, varying the number of retained PCs and discriminant functions and looking for stability in cluster membership. Second, ADMIXTURE (Alexander et al. 2009) was used from  $K = 1$ -6 (inclusive), with 10 replicate runs per value of  $K$  to obtain a per- $K$  coefficient of variation (CV error) to identify the lowest CV error. StructureSelector (Li and Liu 2018) was used on these data to further identify the optimal number of clusters, using the ADMIXTURE output and either the estimated population identifiers from groupings on the PCA, or without different population identifiers.

The PCA, unsupervised DAPC, and ADMIXTURE results were collectively used to define the optimal number of clusters to explain the data. When individuals were assigned to similar clusters concordantly by multiple methods, they were considered assigned to the cluster. If discordant, the result for the individual was considered uncertain, and if the ADMIXTURE ancestry fraction was less than 70% for the major fraction, then the individual was also considered to have uncertain cluster membership, and to have putatively mixed ancestry. A dataset was then generated with only the individuals showing strong cluster membership to the three identified clusters. VCFtools was used to estimate the average  $F_{ST}$  between each cluster.

## 1 *Outlier detection across the coastal-to-interior ecotone*

2 Outlier loci were investigated using the dataset with only the individuals having strong group  
 3 membership (see above) but including all SNP loci (i.e., autosome-only, MAF > 0.05, no LD  
 4 filter). As the focus of the study was the coastal-to-interior ecotone, the three identified clusters  
 5 were defined as either coastal or interior for designating contrasts. First, a supervised DAPC was  
 6 conducted and loading values per locus were obtained; loci within the 99.99<sup>th</sup> percentile of  
 7 loading values were considered to be associated with the regional differences. Loading values  
 8 were plotted in a Manhattan plot in R using fastman (Paria et al. 2022), including only the  
 9 scaffolds with at least 100 SNPs present in the full dataset to remove smaller scaffolds. Second,  
 10 GEMMA (Zhou and Stephens 2012) was used with interior or coastal designations as binary  
 11 values. A kinship matrix was generated among all individuals, and linear models based on the  
 12 binary region definition per locus were conducted. Log-ratio test p-values were plotted in a  
 13 Manhattan plot using fastman as described above, with a significance threshold was determined  
 14 by Bonferroni correction (i.e., 0.05 divided by the number of tests). Third, *pcadapt* (Luu et al.  
 15 2017) was used and a screeplot and scoreplots for the first six principal components (PCs) were  
 16 used to determine whether specific PCs separated the pre-defined clusters. The PC axis that best  
 17 separated the coastal and interior groupings was determined and used as the relevant PC.  
 18 Significance values (p-values) of loci associated with relevant PC produced by *pcadapt* were  
 19 extracted and a multiple test correction was applied using the *p.adjust* function in R using the  
 20 Benjamini-Hochberg correction method to determine significance of individual loci, then  
 21 adjusted p-values were plotted in a Manhattan plot.

22 Comparisons of the different outlier SNP detection methods were conducted by  
 23 identifying outlier loci and regions detected by multiple methods. Proximity to predicted genes  
 24 for top outlier candidates were inspected using the annotation table from NCBI for the UrsArc2.0  
 25 reference genome. Genotypes of top outlier candidates were plotted based on allelic dosage in R.

## 27 **Results**

### 28 *Whole-genome resequencing, genetic sex, and genotyping*

29 The 31 unique grizzly bear samples had an average ( $\pm$  s.d.) number of reads per sample of 121.5  
 30  $\pm$  43.5 M (Additional File S1), or  $15.9 \pm 5.7\times$  coverage, assuming a genome size of 2.3 Gbp.  
 31 Alignment rates against the contig-level genome assembly (GCF\_003584765.1) were on average

82.6  $\pm$  4.9% of total reads. Following GATK genotyping (see *Methods*), 13,867,192 biallelic SNPs were identified. Applying a MAF filter resulted in the retention of 10,044,612 SNPs. Of these, 9,788,257 SNPs were transferred to the UrsArc2.0 genome assembly (see *Methods*). Removal of SNPs that were near indels or that had overall low quality or depth resulted in a minimal number of SNPs removed, and 9,683,999 SNPs were retained. Removal of low or very high depth or low-quality genotypes per individual (see *Methods*) resulted in a significant reduction of SNPs, suggesting the removal of many low-quality genotype calls ( $n = 3,895,954$  SNPs retained). Following these filters, the samples had 9.6  $\pm$  11.0% missing data overall (see Figure S1). The reapplication of a MAF filter resulted in 3,880,487 SNPs being retained in the all SNP dataset, and 327,820 SNPs in the LD-filtered dataset. SNPs on the sex chromosomes or mitochondrial genome were removed from the all SNP and LD-filtered datasets, resulting in the retention of 3,871,837 and 325,946 SNPs, respectively.

Raw read alignments to the UrsArc2.0 assembly were used to determine the genetic sex of individuals by analyzing alignments to the X (121.2 Mbp) and Y (30.9 Mbp) chromosomes (see *Methods*). Clear alignment differences were observed between the sexes; suspected females had very low relative average coverage on the Y-chromosome (coverage of Y/X = 0.04) whereas suspected males had more similar coverage across both chromosomes (coverage of Y/X = 0.26). The determined genetic sexes matched phenotypic sexes provided with the samples, which specified nine females and 22 males.

### *Mitochondrial phylogeny of samples within known clades*

Following the definition of clades identified in Miller et al. (2006), our analysis of previously analyzed haplotypes (Miller et al. 2006) with samples from coastal Alaska (Talbot et al. 2014) and haplotypes identified in our samples resulted in a tree where all clades except for Clade 6 were largely retained (Figure S2). In our results, Clade 6 was split into two paraphyletic clades, with two samples together with Clade 2 and three samples within their own clade. The mitochondrial haplotypes identified in the target mitochondrial region from the samples of the present study ( $n = 4$  unique haplotypes) formed a smaller clade with haplotypes from eastern Russia, coastal Alaska, and central coast BC (Figure S2). This smaller clade clustered within the larger clade with other haplotypes from Alaska and BC, broadly within Clade 3 (Miller et al. 2006; Talbot et al. 2014). Clade 3 has previously been separated into Clade 3a (Eurasia, Alaska,

and Hokkaido Central), 3b (Canada and Hokkaido East), and 3c (Middle East) (de Jong et al. 2023). Clade 3b, as defined in Miller et al. (2006), contains haplotypes 66, 67, and R252 (see Figure 1 of Miller et al. 2006), and our samples clustered with these Clade 3b haplotypes (Figure S2).

### *Population and sample characterization*

Using principal components analysis (PCA) on the LD-filtered genotypes resulted in individuals generally clustering by the geographic location of sampling (Figure 2A). PC1 separated individuals across latitude (southern = negative PC1; northern = positive PC1; percent variance explained, PVE = 6.7%). PC2 separated across longitude, with more western (i.e., coastal) samples in positive PC2 and eastern (interior) samples in negative PC2 (PVE = 4.7%). These clusters included individuals of both sexes and from a wide variety of sampled years (Figure S3), suggesting temporal stability. Some notable exceptions to these trends were observed. Two samples from the southern end of the sampled area (i.e., 121224 and 122177 from south of Bella Coola) clustered closer to the more northern samples. A sample collected from the furthest eastern location in the dataset and one of the furthest south sampling points (i.e., 113981 from the Chilcotin Region) was positioned in the middle of the PCA sample distribution. However, generally the overall groupings corresponded to geographic location of sample collections. Inspecting sex, year of collection, or percentage of missing data did not explain PCA clustering (Figure S3).

Genetic relatedness between samples was estimated and paired relatedness values were found to generally follow a normal distribution with a right tail representing pairs with elevated relatedness (Figure 2B). Four pairs involving five unique individuals exhibited relatedness estimates above the set cutoff (Ritland metric > 0.15; Table S1). Notably, these pairs were comprised of individuals sampled geographically close to each other. The most closely related pair (i.e., 110171 and 67931; Ritland = 0.2448), were males sampled from the western arm of the Nechako reservoir in 2010 and north of Morice River in 1996, respectively (distance between = 61.5 km; Table S1). The other highly related pairs were sampled near Rivers Inlet. Individuals 112146 (female, 2010) and 114228 (male, 2014; Ritland = 0.2149) were sampled in the mountains north of Rivers Inlet approximately 12 km apart, and 102956 (male, 2009), also estimated to be related to this pair (e.g., 114228-102956 Ritland = 0.1816) was sampled

approximately 23 km to the north. As expected, the individuals with elevated relatedness also clustered closely in the PCA (Figure 2A). Individuals 102956 and 114228 were among those samples with higher levels of missing data, but all the others with high relatedness were not within the elevated relatedness pairs (Table S1; Additional File S1), suggesting that the relatedness trend was not caused by missing or low-quality genotypes. Three samples were removed from the dataset to avoid impacts of putative close-kin (i.e., 110171, 114228, 102956), but this did not significantly impact clustering.

### *Defining population clusters*

Population structure was investigated using an unsupervised DAPC and ADMIXTURE ( $n = 28$  individuals; 325,946 SNPs). The DAPC identified a slight elbow in the BIC at  $K = 2$  (Figure S4). Furthermore, DAPC assignment of individuals to clusters was more stable at  $K = 2$  than  $K = 3$ ; when using  $K = 3$ , slight variations in the number of retained PCs in the DAPC resulted in large differences in cluster formation, where the third cluster was frequently comprised of only a few individuals. Therefore,  $K = 2$  was used to assign individuals to clusters by DAPC (Figure S5).

The lowest ADMIXTURE CV error occurred with  $K = 1$  (Figure S6), suggesting that all samples descend from one main ancestry cluster. However, to investigate potential substructure related to the groupings observed in the PCA (see above), and to identify individuals with strong cluster membership to the three PCA groupings observed, we explored  $K = 3$ . This analysis separated samples largely into the three groupings observed in the PCA (coastal south, coastal north, and interior; Figure 3; Figure S7B).

Using ADMIXTURE  $K = 3$  and considering only individuals with ancestry fractions greater than 70% as well as consistent groupings by both ADMIXTURE and DAPC, the coastal south, coastal north, and interior clusters contained six individuals each ( $n = 18$  total; Figure 3; Figure S7B). Individuals 122177 and 72 showed discrepancies between the methods in cluster assignment, and so even though they had >70% ancestry fractions, they were not retained in the interior grouping.

Although sample sizes were generally low, to understand the extent of genetic differentiation between these groupings, average  $F_{ST}$  was evaluated between groups. The differentiation analysis was in concordance with the PCA in that the greatest difference was

observed by latitude, with north coastal and south coastal  $F_{ST} = 0.036$ , relative to north coastal and interior being  $F_{ST} = 0.015$  (Table S2).

#### *Outlier detection and genomic characterization*

Outlier identification was conducted through supervised analyses of coastal (coastal north and coastal south;  $n = 12$ ) and interior ( $n = 6$ ) groupings using several methods. The supervised DAPC approach used 3,858,384 loci and resulted in discrete separation between the groups across discriminant function (DF) 1 (Figure S8). The 0.01% of loci with top loading values were identified ( $n = 386$  SNPs), and these had loadings per allele ranging from  $2.15E-6$  to  $3.40E-6$  (median loading contribution of all loci =  $5.4E-8$ ; median of top 99.99<sup>th</sup> percentile loci =  $2.4E-6$ ; i.e., 44.0x higher median in the outliers). These top outliers were found throughout the genome, including regions with multiple SNPs observed in peaks (Figure 4A). The pcadapt approach used 3,706,201 loci and found PC1 to explain the latitudinal separation, as was observed for PC1 in the main PCA (see above), and PC2 to explain the coastal/interior separation (Figure S9). Significant outlier SNPs (Benjamini-Hochberg corrected  $p < 0.01$ ) were identified that were associated with PC2 ( $n = 4,391$  SNPs). These loci were also found throughout the genome, with some having very low p-values (Figure 4B). The GEMMA approach analyzed 703,001 loci and found 24 SNPs to be significantly associated with the coastal/interior separation (Bonferroni-adjusted  $p \leq 0.05$ ). These were found on seven different scaffolds (Figure 4C).

Significant outlier SNPs found by multiple methods were then inspected, either by identifying the exact SNP by more than one approach or by identifying other SNPs in the same genomic region. There were 128 outlier SNPs identified by both DAPC and pcadapt (i.e., 33% of the DAPC outliers and 2.9% of pcadapt outliers). Six genomic scaffolds each contain at least five of these 128 common outliers (Table 1; Additional File S2). This includes 83 SNPs on scaffold NW\_026622875 between 19.86-19.97 Mbp (Table 1). Scaffold NW\_026622863 has multiple regions with common outliers including near 47.88 Mbp and 56.63 Mbp. There were 21 SNPs identified by both pcadapt and GEMMA, with 11 of these being on NW\_025522863 between 47.75-47.82 Mbp (Table 1). Although no SNPs were identified by all three methods, there were consistently identified regions shared by all three approaches, including most notably the region around 47.8 Mbp of NW\_026622863 (Table 1). A full list of SNPs associated as identified by each method, and shared loci between methods, is present in Additional File S2.

The consistently identified outlier genomic regions were inspected for gene content using the gene annotation for UrsArc2.0 (accessed June 11<sup>th</sup>, 2025). The region of interest identified by all three methods, NW\_025522863 between 47,746,156-47,879,257 bp is a 133,101 bp region that contains 2, 96, and 11 significant SNP outliers for DAPC, pcadapt, and GEMMA, respectively (Figure 5). This region contains seven predicted genes (see Table 1). Most notably, at 47,888,920 – 47,897,638 (9.7 kb downstream) is the annotated gene *creatine kinase, M-type* (CKM; Table 1). Several top significant SNPs in this region that were found by multiple outlier detection methods were plotted for allelic dosage (Figure 6), including pcadapt and GEMMA outliers, including the most significant pcadapt outliers for PC2, a SNP at 47.746 Mbp and one at 47.822 Mbp. These two SNPs show similar genotypic patterns where all coastal individuals are homozygous for the reference allele and most inland individuals are heterozygous (with one being homozygous alternate). Shared DAPC and pcadapt significant outliers were also identified in this region including a SNP at 47.870 Mbp and one at 47.880 Mbp (Figure 5; Figure 6).

The other region on scaffold NW\_025522863 identified by both pcadapt and DAPC at 56.63 Mbp is near a zinc finger gene (*zinc finger protein OZF-like*), and multiple other zinc finger genes upstream (i.e., 11 unique predicted zinc finger genes from 56.40-56.62 Mbp). The region with 83 shared SNPs by both pcadapt and DAPC at 19.86-19.97 Mbp of NW\_026622875 is near a predicted gene annotated as *solute carrier family 9 member A9* (Table 1).

## Discussion

Genomic investigations of SNP variants putatively under selection have identified locally adapted populations across small geographic areas (Richardson et al. 2014), in wide-ranging highly-mobile species (Schweizer et al. 2016b), and in the presence of gene flow (Tigano and Friesen 2016). Our findings of substructure within the BC sampling range and presence of putative outlier loci across the Coastal Mountain Range (CMR) indicates the potential for local adaptation in grizzly bears in BC. Additionally, the dataset presented here provides a valuable new genomic resource for grizzly bears that fills a sampling gap through western Canada's Central Coast in the recently published global analysis of whole-genome resequencing of brown bears (de Jong et al. 2023).

# Population genomic trends in BC grizzly bears

The present data showed clustering of grizzlies by the geographic region of sampling (i.e., coastal north, coastal south, and interior), with some individuals having unclear or mixed genetic backgrounds. This clustering was observed after removing individuals with elevated estimated genetic relatedness, which was important given the potential impact of closely related individuals on the tools applied here (i.e., PCA, ADMIXTURE; Patterson et al. 2006; Anderson and Dunham 2008; Elhaik 2022; Yao and Ochoa 2023). The greatest genetic differentiation (i.e., substructuring) in the dataset was based on latitude, not based on coastal or interior delineations. Pairs of grizzlies were found with elevated relatedness, and these pairs were sampled in geographically proximal locations but in some cases many years apart. Furthermore, increased genetic similarity in proximal geographic areas was independent of year sampled, suggesting that these genetic neighbourhoods have persisted over time, even though there are some individuals present with putatively mixed genetic backgrounds among the clusters, and therefore some gene flow could be expected.

Although there was substantial genome-wide  $F_{ST}$  observed between the three genetic clusters (most notably between the southern and northern clusters), the ADMIXTURE analysis found  $K = 1$  as the best model to explain the data. This observation suggests that all the clusters originate from the same overall genetic ancestry but may have been impacted by genetic drift relatively recently, resulting in allele frequency changes in the individual clusters, with some gene flow likely (based on the presence of individuals with mixed ancestry fractions with a  $K = 3$  ADMIXTURE model). Isolation-by-distance (IBD) can challenge the analytic approaches used here; ADMIXTURE assumes random mating, but IBD can violate this assumption (Lawson et al. 2018), which in some cases can lead to an overestimation of  $K$  (Frantz et al. 2009). However, in the present study, spatial separation does not appear to have overestimated  $K$ . It was valuable here to use these multiple different approaches to understand the genetic trends in the study (Lawson et al. 2018).

The genetic differentiation among geographic regions was contrasted by the presence of individuals with mixed ancestry fractions (when using  $K = 3$ ) that were not clearly assigned to any one of the three identified clusters. The area around Bella Coola had several sampled individuals with putatively mixed genetic backgrounds that were therefore not included in the outlier identification. The Bella Coola Valley bisects the CMR, and therefore may provide a



corridor for access to the coastal area. Similarly, an individual with a putatively mixed genetic background was observed on the edge of the designated coastal region near the Kitimat and Kemano valleys connecting the interior to the coastal region. Grizzlies were also observed further into the interior region with mixed  $K = 3$  ancestry fractions. The movement of interior grizzly bears through the Bella Coola valley to access salmon has long been known by the Nuxalk Nation (Jason Moody, *pers. comm.*). If valleys traversing the CMR are used as corridors, this provides opportunities for connectivity maintenance for management. These results would benefit from additional samples to improve our understanding of the extent of connectivity and movement during reproductive seasons. Although the present study was limited to the tissues that were available, a more continuous and expanded sampling strategy could improve our understanding of connectivity among these regions, including the potential presence of any significant barriers. Henson et al. (2021) also identified three separate groupings (described as STRUCTURE populations) in BC, and these also followed a trend of separating by geography with some overlap and the presence of putative migrant individuals in the different areas.

Bears are expected to move at different times of year, for example to access salmon in the fall, and for mating in the spring. This natural movement process would provide an important link between coastal and interior habitats. Interestingly, Bella Coola and Kitimat/Kemano also align with two of the most important and widely used eulachon *Thaleichthys pacificus* grease trails used and maintained to connect Indigenous communities in trade and other processes over millennia (Harrington 1953). Such convergence in use between species emphasizes how humans and bears can be similarly shaped by landscapes (Henson et al. 2021), and the long-term importance of these valleys for bears and people to access coastal resources.

#### *Local adaptation and putative outlier loci across the coastal-to-interior ecotone*

Although latitude was the most significant explanatory factor for genetic variation, by contrasting interior and coastal regions, outlier loci were detected that provide candidate genomic regions potentially associated with phenotypic differences observed between the coastal and interior regions. Genomic regions with clusters of SNPs identified by multiple approaches were of particular interest, most notably the region from 47.75-47.88 Mbp of scaffold NW\_026622863.1. This region is 9.7 kbp upstream from the single copy gene *creatine-kinase, m-type* (CKM). This gene is expressed predominantly in muscle and heart tissues of grizzlies of

both sexes, as observed based on exon expression profiles in NCBI (PRJNA413091; Jansen et al. 2019). CKM is the muscle type of creatine kinases, and is involved in energy homeostasis, catalyzing the reversible transfer of phosphate from ATP to creatine to produce phosphocreatine (Abnous and Storey 2007), a temporary energy storage in muscle (Whiteman et al. 2017). During food shortages, polar bears reduce activity and use stored energy (e.g., in spring following winter food deprivation); alongside the reduced muscle protein concentration and increased water content that occurs during atrophy, reduced expression of *ckm* mRNA is also observed (Whiteman et al. 2017). In the ground squirrel *Spermophilus richardsonii*, CK activity and protein levels are reduced during hibernation, and *ckm* mRNA expression is reduced by 70% (Abnous and Storey 2007). The physiological role of CKM in energy metabolism associated with intermittent food availability and stores makes this gene an interesting candidate given its proximity to the most consistently identified outlier region between coastal and interior grizzly bears here. This genomic region, and other candidate regions, including the second peak further downstream on the same scaffold that is within a region replete with zinc finger protein-encoding genes (often involved in transcription regulation; Cassandri et al. 2017), merit further investigation in future studies in terms of their potential roles in the differential phenotypes observed between coastal and interior grizzly bears.

Known phenotypic differences exist between coastal and interior BC grizzly bears, and these different regions across the coastal-interior ecotone have significant ecosystem differences to which the residents would be exposed. Key phenotypic and environment differences have been documented between larger, salmon- and intertidal-foraging coastal bears and smaller, interior bears in terms of morphology (Rausch 1963; Kurtén 1973; Paetkau et al. 1998), resource use (Adams et al. 2017), and potential pathogen pressure (Catalano et al. 2015; Robbins et al. 2018). In coastal bears, adaptations for enhanced growth may have arisen in response to their greater access to, consumption of, and size-mediated competition over salmon (Gende and Quinn 2004; Robbins et al. 2007; Service et al. 2019). In contrast, growth inhibition in interior bears would be advantageous for regulating body mass, given local intermittent access to high-protein resources (Felicetti et al. 2003). The differential pathogen pressures presented by either primarily cervid- or salmon-associated meat resources in interior and coastal areas, respectively, could also result in immune-related adaptations in each area (Catalano et al. 2015; Robbins et al. 2018). The outlier loci identified in the present study may be related to these phenotypic and ecotypic

1 differences, including resource niche differentiation of coastal and interior grizzly bears.  
2 However, they are likely only part of a complex suite of polygenic and epigenetic differences  
3 that interact with diet-induced patterns of phenotypic plasticity.

4 Increased understanding of the underlying environmental factors that drive local  
5 adaptation can help to identify loci associated with local adaptation (Booker 2024). Improved  
6 characterization and analytic use of the drivers of the main selective forces on grizzlies across  
7 the ecotone may therefore improve our ability to detect loci associated with this selection.  
8 Without an exact characterization of the environment that each grizzly experienced for extended  
9 periods of time, the present study relied upon sampling location to contrast the different  
10 subpopulations (with an attempt to exclude putative migrants). If an environmental variable  
11 suspected of being a driver of selection across the ecotone was known, and the per-grizzly value  
12 of this variable was obtained and usable in the association analysis, this could improve resolution  
13 of the genomic associations to the ecotone. However, it is not clear whether this exact  
14 specification of a continuous variable per individual would be possible for grizzlies, considering  
15 their wide-ranging habitats, and is not possible with the current dataset, and therefore we relied  
16 upon general geographic groupings that were classified as either coastal or interior.  
17 Understanding selection can also be improved by considering dispersal patterns alongside  
18 environmental variation (Booker 2024). Importantly, when selective pressures and population  
19 structure are co-autocorrelated over geographic areas (as could be expected in grizzlies across  
20 the CMR ecotone), local adaptation can be strong (depending on gene flow and strength of  
21 selection), but it may also be more difficult to characterize (Booker 2024). The present study  
22 gives initial insights into this question across the BC CMR ecotone in grizzly bears.

23 Increased sample sizes from each region of the study would improve resolution and  
24 reduce the potential for false positive outlier detection. In addition to the relatively low sample  
25 size, another potential shortcoming of our approach is the grouping of southern and northern  
26 coastal samples together to compare with the interior samples (that are more genetically similar  
27 to the northern coastal cluster than the southern coastal). This approach assumes that the two  
28 coastal regions, although they have the greatest differentiation in the dataset, will have had  
29 parallel adaptations or consistent genotypic variation across the ecotone. There is also a  
30 possibility of confounding latitudinal variation with ecotone-related variation, although  
31 inspections of individual loci for top outliers show consistent genotypes in both southern and

northern coastal areas contrasted with the interior. Removing the southern coastal samples and only analyzing the coastal-to-interior contrast at a similar latitude would reduce the sample size by a third, and therefore be expected to significantly reduce detection power. Additionally, the north-to-south variation was mainly captured by a separate axis of variation in the *pcadapt* analysis. In any case, the findings of regions putatively linked to the alternate sides of the coastal-to-interior ecotone are valuable for future studies but should also be considered as initial evidence for involvement and not definitive. Further evaluating the associations of these regions to segregating phenotypic variation across the coastal-to-interior ecotone will be important in future work.

### *Management implications*

Our results indicating population substructure and potential local adaptation have implications that can be considered for management applications. For example, these results indicate gene flow among provincially-designated Grizzly Bear Population Units (GBPUs), as well as the potential for locally adapted regional groups. Individuals from the genetically continuous interior group span multiple current GBPUs (i.e., Tweedsmuir and Bulkley GBPUs; Fig. 1), which emphasizes the need to maintain connectivity among these management units. Furthermore, the GBPU system may not adequately describe, integrate, and protect corridors bisecting the CMR to connect coastal and interior groups. Although the restricted geographic sampling in the present study limits inference regarding the spatial designation of coastal GBPUs, evidence for coastal latitudinal genetic differentiation was observed. Although it was not formally evaluated here given the focus on the coast-to-interior ecotone, the present data could also be investigated for outliers between the two coastal regions, albeit with a reduced sample size to the present analysis.

Interior grizzlies have adapted to more extreme environmental conditions found in continental climates and have regulations on body size presumably related to intermittent resources (Robbins et al. 2007). Coastal individuals may lack such adaptations to maintain body size, which could pose a risk under a future defined by ever decreasing populations of salmon (Price et al. 2008). With increased variation in environmental conditions expected, adaptations for these fluctuating conditions may not be present in the coastal group (Felicetti et al. 2003). The immunological capacities may also differ between regions; for example, interior individuals

may lack immunity to pathogens present in the coastal environment presented by intertidal and marine protein resources (Catalano et al. 2015; Robbins et al. 2018). The possibility of these vulnerabilities and the demonstrated susceptibility elsewhere of locally adapted populations to environmental stressors (Valladares et al. 2014; Anderson and Wadgymar 2020) highlights the value of future research investigating local adaptation in these regions, as well as the frequency and extent of gene flow. Our findings also support cautious management practices designed to preserve gene flow between coastal and interior groups and protect salmon and coastal habitats as resources linked to the evolutionary history and future productivity of potentially uniquely adapted coastal grizzly bears.

## Conclusions

Here we used whole-genome resequencing to improve our understanding of the genetic differences across the coastal-to-interior ecotone in BC, and in doing so identified three distinct subpopulation clusters (i.e., north coastal, south coastal, and interior). By inspecting grizzlies identified as predominantly belonging to each of the three subpopulations, we identified segregating genetic variants and associated genomic regions and candidate genes between the coastal and interior regions and signatures of potential local adaptation. With continued environmental or resource changes in each region, local adaptation will be important to consider in terms of resiliency of grizzly bears from different geographic regions. These results suggest that it will be important for management to consider both the connectivity corridors between regions, but also the potential for locally adapted and unique subpopulations depending on the geographic region.

## Data Availability

Raw sequence data have been uploaded to NCBI's short read archive (SRA) under BioProject PRJNA1204358 within accessions SAMN46039649-SAMN46039679. Datashare agreements with the Province of British Columbia restrict the sharing of precise locations where samples were obtained. VCF files containing sample multi-locus genotypes are available through the G3 FigShare portal (<https://doi.org/10.25387/g3.30090427>).

The following code repositories support this project:

1 Manuscript code repository: [https://github.com/bensutherland/ms\\_grizzly\\_popgen](https://github.com/bensutherland/ms_grizzly_popgen)

2 Population genetics analysis: [https://github.com/bensutherland/simple\\_pop\\_stats](https://github.com/bensutherland/simple_pop_stats)

3 Additional bioinformatics functions: <https://github.com/enormandeau/Scripts>

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## 18 **Competing Interests**

19 BJGS is affiliated with Sutherland Bioinformatics. The author has no competing financial  
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## 23 **Author Contributions**

24 Lauren Henson, Jason Moody, and Chris Darimont contributed to hypothesis formation. Lauren  
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26 Stronen contributed to data preparation and analysis. Paul Paquet, Jason Moody, Bridgett  
27 vonHoldt, Chris Darimont, Ben Koop, and Astrid Vik Stronen provided conceptual and  
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## References

- Abnous, K., and K.B. Storey, 2007 Regulation of skeletal muscle creatine kinase from a hibernating mammal. *Archives of Biochemistry and Biophysics* 467 (1):10-19.
- Adams, M.S., T. Levi, M. Bourbonnais, C.N. Service, K. Artelle *et al.*, 2024 Human disturbance in riparian areas disrupts predator-prey interactions between grizzly bears and salmon. *Ecology and Evolution* 14 (3):e11058.
- Adams, M.S., C.N. Service, A. Bateman, M. Bourbonnais, K.A. Artelle *et al.*, 2017 Intrapopulation diversity in isotopic niche over landscapes: spatial patterns inform conservation of bear-salmon systems. *Ecosphere* 8 (6):e01843.
- Alexander, D.H., J. Novembre, and K. Lange, 2009 Fast model-based estimation of ancestry in unrelated individuals. *Genome Research* 19 (9):1655-1664.
- Alvarado, A.H., T.L. Fuller, and T.B. Smith, 2014 Integrative tracking methods elucidate the evolutionary dynamics of a migratory divide. *Ecology and Evolution* 4 (17):3456-3469.
- Anderson, E.C., and K.K. Dunham, 2008 The influence of family groups on inferences made with the program Structure. *Molecular Ecology Resources* 8 (6):1219-1229.
- Anderson, J.T., and S.M. Wadgymar, 2020 Climate change disrupts local adaptation and favours upslope migration. *Ecology Letters* 23 (1):181-192.
- Armstrong, E.E., B.W. Perry, Y. Huang, K.V. Garimella, H.T. Jansen *et al.*, 2022 A beary good genome: haplotype-resolved, chromosome-level assembly of the brown bear (*Ursus arctos*). *Genome Biology and Evolution* 14 (9).
- Barry, P.D., D.A. Tallmon, N.S. Howe, D.S. Baetscher, K.L. D'Amelio *et al.*, 2024 A major effect locus involved in migration timing is shared by pink and sockeye salmon. *bioRxiv*:2024.2003.2030.587279.
- Benestan, L., J.S. Moore, B.J.G. Sutherland, J. Le Luyer, H. Maaroufi *et al.*, 2017 Sex matters in massive parallel sequencing: Evidence for biases in genetic parameter estimation and investigation of sex determination systems. *Molecular Ecology* 26 (24):6767-6783.
- Bennuah, S.Y., T. Wang, and S.N. Aitken, 2004 Genetic analysis of the *Picea sitchensis* × *glauca* introgression zone in British Columbia. *Forest Ecology and Management* 197 (1):65-77.
- Blanquart, F., O. Kaltz, S.L. Nuismer, and S. Gandon, 2013 A practical guide to measuring local adaptation. *Ecology Letters* 16 (9):1195-1205.
- Booker, T.R., 2024 The structure of the environment influences the patterns and genetics of local adaptation. *Evolution Letters* 8 (6):787-798.
- Broad Institute, 2024 Picard toolkit, GitHub Respository.
- Bryan, H.M., C.T. Darimont, P.C. Paquet, K.E. Wynne-Edwards, and J.E.G. Smits, 2013 Stress and reproductive hormones in grizzly bears reflect nutritional benefits and social consequences of a salmon foraging niche. *PLOS ONE* 8 (11):e80537.
- Burford, M.O., J. Scarpa, B.J. Cook, and M.P. Hare, 2014 Local adaptation of a marine invertebrate with a high dispersal potential: evidence from a reciprocal transplant experiment of the eastern oyster *Crassostrea virginica*. *Marine Ecology Progress Series* 505:161-175.
- Cahill, J.A., P.D. Heintzman, K. Harris, M.D. Teasdale, J. Kapp *et al.*, 2018 Genomic Evidence of Widespread Admixture from Polar Bears into Brown Bears during the Last Ice Age. *Mol Biol Evol* 35 (5):1120-1129.

- 1 Cahill, J.A., I. Stirling, L. Kistler, R. Salamzade, E. Ersmark *et al.*, 2015 Genomic evidence of  
2 geographically widespread effect of gene flow from polar bears into brown bears.  
3 *Molecular Ecology* 24 (6):1205-1217.
- 4 Cassandri, M., A. Smirnov, F. Novelli, C. Pitolli, M. Agostini *et al.*, 2017 Zinc-finger proteins in  
5 health and disease. *Cell Death Discovery* 3 (1):17071.
- 6 Catalano, S., M. Lejeune, P. Tizzani, G.G. Verocai, H. Schwantje *et al.*, 2015 Helminths of  
7 grizzly bears (*Ursus arctos*) and American black bears (*Ursus americanus*) in Alberta  
8 and British Columbia, Canada. *Canadian Journal of Zoology* 93 (10):765-772.
- 9 Dalongeville, A., L. Benestan, D. Mouillot, S. Lobreaux, and S. Manel, 2018 Combining six  
10 genome scan methods to detect candidate genes to salinity in the Mediterranean striped  
11 red mullet (*Mullus surmuletus*). *BMC Genomics* 19 (1):217.
- 12 Danecek, P., A. Auton, G. Abecasis, C.A. Albers, E. Banks *et al.*, 2011 The variant call format  
13 and VCFtools. *Bioinformatics* 27 (15):2156-2158.
- 14 Danecek, P., J.K. Bonfield, J. Liddle, J. Marshall, V. Ohan *et al.*, 2021 Twelve years of  
15 SAMtools and BCFtools. *GigaScience* 10 (2).
- 16 Darimont, C.T., T.E. Reimchen, and P.C. Paquet, 2003 Foraging behaviour by gray wolves on  
17 salmon streams in coastal British Columbia. *Canadian Journal of Zoology* 81 (2):349-  
18 353.
- 19 de Jong, M.J., A. Niamir, M. Wolf, A.C. Kitchener, N. Lecomte *et al.*, 2023 Range-wide whole-  
20 genome resequencing of the brown bear reveals drivers of intraspecies divergence.  
21 *Communications Biology* 6 (1):153.
- 22 Ekar, J.M., D.K. Price, M.A. Johnson, and E.A. Stacy, 2019 Varieties of the highly dispersible  
23 and hypervariable tree, *Metrosideros polymorpha*, differ in response to mechanical stress  
24 and light across a sharp ecotone. *American Journal of Botany* 106 (8):1106-1115.
- 25 Elhaik, E., 2022 Principal component analyses (PCA)-based findings in population genetic  
26 studies are highly biased and must be reevaluated. *Scientific Reports* 12 (1):14683.
- 27 Felicetti, Laura A., Charles T. Robbins, and Lisa A. Shipley, 2003 Dietary protein content alters  
28 energy expenditure and composition of the mass gain in grizzly bears (*Ursus arctos*  
29 *horribilis*). *Physiological and Biochemical Zoology* 76 (2):256-261.
- 30 Frantz, A.C., S. Cellina, A. Krier, L. Schley, and T. Burke, 2009 Using spatial Bayesian methods  
31 to determine the genetic structure of a continuously distributed population: clusters or  
32 isolation by distance? *Journal of Applied Ecology* 46 (2):493-505.
- 33 Funk, W.C., B.R. Forester, S.J. Converse, C. Darst, and S. Morey, 2019 Improving conservation  
34 policy with genomics: a guide to integrating adaptive potential into U.S. Endangered  
35 Species Act decisions for conservation practitioners and geneticists. *Conservation*  
36 *Genetics* 20 (1):115-134.
- 37 Gende, S.M., and T.P. Quinn, 2004 The relative importance of prey density and social  
38 dominance in determining energy intake by bears feeding on Pacific salmon. *Canadian*  
39 *Journal of Zoology* 82 (1):75-85.
- 40 Gruber, B., P.J. Unmack, O.F. Berry, and A. Georges, 2018 dartr: An r package to facilitate  
41 analysis of SNP data generated from reduced representation genome sequencing.  
42 *Molecular Ecology Resources* 18 (3):691-699.
- 43 Halbritter, A.H., S. Fior, I. Keller, R. Billeter, P.J. Edwards *et al.*, 2018 Trait differentiation and  
44 adaptation of plants along elevation gradients. *Journal of Evolutionary Biology* 31  
45 (6):784-800.



- 1 Harrington, L., 1953 On the trail of the Candlefish, pp. 40-44 in *The Beaver*. The Hudson's Bay  
2 Company and Canada's National History Society.
- 3 Henson, L.H., N. Balkenhol, R. Gustas, M. Adams, J. Walkus *et al.*, 2021 Convergent  
4 geographic patterns between grizzly bear population genetic structure and Indigenous  
5 language groups in coastal British Columbia, Canada. *Ecology and Society* 26 (3).
- 6 Hilderbrand, G.V., S.D. Farley, C.T. Robbins, T.A. Hanley, K. Titus *et al.*, 1996 Use of stable  
7 isotopes to determine diets of living and extinct bears. *Canadian Journal of Zoology* 74  
8 (11):2080-2088.
- 9 Hilderbrand, G.V., C.C. Schwartz, C.T. Robbins, M.E. Jacoby, T.A. Hanley *et al.*, 1999 The  
10 importance of meat, particularly salmon, to body size, population productivity, and  
11 conservation of North American brown bears. *Canadian Journal of Zoology* 77 (1):132-  
12 138.
- 13 Hoban, S., J.L. Kelley, K.E. Lotterhos, M.F. Antolin, G. Bradburd *et al.*, 2016 Finding the  
14 genomic basis of local adaptation: pitfalls, practical solutions, and future directions. *The*  
15 *American Naturalist* 188 (4):379-397.
- 16 Hoeksema, Jason D., and Samantha E. Forde, 2008 A meta-analysis of factors affecting local  
17 adaptation between interacting species. *The American Naturalist* 171 (3):275-290.
- 18 Jansen, H.T., S. Trojahn, M.W. Saxton, C.R. Quackenbush, B.D. Evans Hutzenbiler *et al.*, 2019  
19 Hibernation induces widespread transcriptional remodeling in metabolic tissues of the  
20 grizzly bear. *Communications Biology* 2 (1):336.
- 21 Jombart, T., and I. Ahmed, 2011 adegenet 1.3-1: new tools for the analysis of genome-wide SNP  
22 data. *Bioinformatics* 27 (21):3070-3071.
- 23 Katoh, K., J. Rozewicki, and K.D. Yamada, 2017 MAFFT online service: multiple sequence  
24 alignment, interactive sequence choice and visualization. *Briefings in Bioinformatics* 20  
25 (4):1160-1166.
- 26 Kawecki, T.J., and D. Ebert, 2004 Conceptual issues in local adaptation. *Ecology Letters* 7  
27 (12):1225-1241.
- 28 Knaus, B.J., and N.J. Grünwald, 2017 vcfr: a package to manipulate and visualize variant call  
29 format data in R. *Molecular Ecology Resources* 17 (1):44-53.
- 30 Kumar, S., G. Stecher, M. Li, C. Knyaz, and K. Tamura, 2018 MEGA X: molecular evolutionary  
31 genetics analysis across computing platforms. *Molecular Biology and Evolution* 35  
32 (6):1547-1549.
- 33 Kurtén, B., 1973 Transberingian relationships of *Ursus arctos linne* (brown and grizzly bears).  
34 Societas Scientiarum Fennica, c, Helsinki .:
- 35 Lawson, D.J., L. van Dorp, and D. Falush, 2018 A tutorial on how not to over-interpret  
36 STRUCTURE and ADMIXTURE bar plots. *Nature Communications* 9 (1):3258.
- 37 Li, H., 2013 Aligning sequence reads, clone sequences and assembly contigs with BWA-MEM,  
38 pp. arXiv:1303.3997.
- 39 Li, Y.-L., and J.-X. Liu, 2018 StructureSelector: A web-based software to select and visualize  
40 the optimal number of clusters using multiple methods. *Molecular Ecology Resources* 18  
41 (1):176-177.
- 42 Lotterhos, K.E., and M.C. Whitlock, 2015 The relative power of genome scans to detect local  
43 adaptation depends on sampling design and statistical method. *Molecular Ecology* 24  
44 (5):1031-1046.

- 1 Miller, C.R., L.P. Waits, and P. Joyce, 2006 Phylogeography and mitochondrial diversity of  
2 extirpated brown bear (*Ursus arctos*) populations in the contiguous United States and  
3 Mexico. *Molecular Ecology* 15 (14):4477-4485.
- 4 Miller, J.M., R.M. Malenfant, L.R. Rivkin, T.C. Atwood, S. Baryluk *et al.*, 2024 Development of  
5 an 8K SNP chip to assess adaptive diversity and hybridization in polar bears.  
6 *Conservation Genetics Resources* 16 (3):237-249.
- 7 Mowat, G., and D.C. Heard, 2006 Major components of grizzly bear diet across North America.  
8 *Canadian Journal of Zoology* 84 (3):473-489.
- 9 Muñoz-Fuentes, V., C.T. Darimont, R.K. Wayne, P.C. Paquet, and J.A. Leonard, 2009  
10 Ecological factors drive differentiation in wolves from British Columbia. *Journal of*  
11 *Biogeography* 36 (8):1516-1531.
- 12 Normandeau, E., M. de Ronne, and D. Torkamaneh, 2023 SNPLift: Fast and accurate conversion  
13 of genetic variant coordinates across genome assemblies.  
14 *bioRxiv*:2023.2006.2013.544861.
- 15 Paetkau, D., G.F. Shields, and C. Strobeck, 1998 Gene flow between insular, coastal and interior  
16 populations of brown bears in Alaska. *Mol Ecol* 7 (10):1283-1292.
- 17 Paria, S.S., S.R. Rahman, and K. Adhikari, 2022 fastman: A fast algorithm for visualizing  
18 GWAS results using Manhattan and Q-Q plots. *bioRxiv*:2022.2004.2019.488738.
- 19 Patterson, N., A.L. Price, and D. Reich, 2006 Population structure and eigenanalysis. *PLOS*  
20 *Genetics* 2 (12):e190.
- 21 Perrier, C., Q. Rougemont, and A. Charmanier, 2020 Demographic history and genomics of  
22 local adaptation in blue tit populations. *Evolutionary Applications* 13 (6):1145-1165.
- 23 Pew, J., P.H. Muir, J. Wang, and T.R. Frasier, 2015 related: an R package for analysing pairwise  
24 relatedness from codominant molecular markers. *Mol Ecol Resour* 15 (3):557-561.
- 25 Price, M.H.H., C.T. Darimont, N.F. Temple, and S.M. MacDuffee, 2008 Ghost runs:  
26 management and status assessment of Pacific salmon (*Oncorhynchus* spp.) returning to  
27 British Columbia's central and north coasts. *Canadian Journal of Fisheries and Aquatic*  
28 *Sciences* 65 (12):2712-2718.
- 29 Province of British Columbia, 2012 British Columbia grizzly bear population estimate for 2012,  
30 edited by M.o.F.L.a.N.R. Operations. Province of B.C., Victoria, BC.
- 31 Quinlan, A.R., and I.M. Hall, 2010 BEDTools: a flexible suite of utilities for comparing genomic  
32 features. *Bioinformatics* 26 (6):841-842.
- 33 Quinn, T.P., 2018 *The behavior and ecology of Pacific salmon and trout*: University of  
34 Washington Press.
- 35 R Core Team, 2025 R: A Language and Environment for Statistical Computing. R Foundation  
36 for Statistical Computing, Vienna, Austria.
- 37 Rausch, R.L., 1963 Geographic variation in size in North American brown bears, *Ursus arctos*  
38 L., as indicated by condylobasal length. *Canadian Journal of Zoology* 41 (1):33-45.
- 39 Richardson, J.L., M.C. Urban, D.I. Bolnick, and D.K. Skelly, 2014 Microgeographic adaptation  
40 and the spatial scale of evolution. *Trends in Ecology & Evolution* 29 (3):165-176.
- 41 Ritland, K., 1996 Estimators for pairwise relatedness and individual inbreeding coefficients.  
42 *Genetics Research* 67 (2):175-185.
- 43 Robbins, C.T., J.K. Fortin, K.D. Rode, S.D. Farley, L.A. Shipley *et al.*, 2007 Optimizing protein  
44 intake as a foraging strategy to maximize mass gain in an omnivore. *Oikos* 116  
45 (10):1675-1682.

- 1 Robbins, C.T., N.L. Woodford, G. Goolsby Clyde, C. Minor, O.L. Nelson *et al.*, 2018 Salmon  
2 poisoning disease in grizzly bears with population recovery implications. *The Journal of*  
3 *Wildlife Management* 82 (7):1396-1402.
- 4 Rosenblum, Erica B., 2006 Convergent evolution and divergent selection: lizards at the White  
5 Sands ecotone. *The American Naturalist* 167 (1):1-15.
- 6 Ruegg, K., 2008 Genetic, morphological, and ecological characterization of a hybrid zone that  
7 spans a migratory divide. *Evolution* 62 (2):452-466.
- 8 Sanford, E., and M.W. Kelly, 2011 Local adaptation in marine invertebrates. *Annual Review of*  
9 *Marine Science* 3 (Volume 3, 2011):509-535.
- 10 Schweizer, R.M., J. Robinson, R. Harrigan, P. Silva, M. Galverni *et al.*, 2016a Targeted capture  
11 and resequencing of 1040 genes reveal environmentally driven functional variation in  
12 grey wolves. *Mol Ecol* 25 (1):357-379.
- 13 Schweizer, R.M., B.M. vonHoldt, R. Harrigan, J.C. Knowles, M. Musiani *et al.*, 2016b Genetic  
14 subdivision and candidate genes under selection in North American grey wolves.  
15 *Molecular Ecology* 25 (1):380-402.
- 16 Service, C.N., A.W. Bateman, M.S. Adams, K.A. Artelle, T.E. Reimchen *et al.*, 2019 Salmonid  
17 species diversity predicts salmon consumption by terrestrial wildlife. *Journal of Animal*  
18 *Ecology* 88 (3):392-404.
- 19 Talbot, S.L., S.A. Sonsthagen, G.K. Sage, S.D. Farley, N.G. Dawson *et al.*, 2014 Are island  
20 brown bears isolated? Insularity and gene flow among coastal populations in southeast  
21 Alaska, edited by U.S.G. Survey. Alaska Science Center, Anchorage, AK.
- 22 Tamura, K., and M. Nei, 1993 Estimation of the number of nucleotide substitutions in the control  
23 region of mitochondrial DNA in humans and chimpanzees. *Molecular Biology and*  
24 *Evolution* 10 (3):512-526.
- 25 Taylor, G.A., H. Kirk, L. Coombe, S.D. Jackman, J. Chu *et al.*, 2018 The genome of the North  
26 American brown bear or grizzly: *Ursus arctos* ssp. *horribilis*. *Genes* 9 (12).
- 27 Tigano, A., and V.L. Friesen, 2016 Genomics of local adaptation with gene flow. *Molecular*  
28 *Ecology* 25 (10):2144-2164.
- 29 Tigano, A., A.J. Shultz, S.V. Edwards, G.J. Robertson, and V.L. Friesen, 2017 Outlier analyses  
30 to test for local adaptation to breeding grounds in a migratory arctic seabird. *Ecology and*  
31 *Evolution* 7 (7):2370-2381.
- 32 Valladares, F., S. Matesanz, F. Guilhaumon, M.B. Araújo, L. Balaguer *et al.*, 2014 The effects of  
33 phenotypic plasticity and local adaptation on forecasts of species range shifts under  
34 climate change. *Ecology Letters* 17 (11):1351-1364.
- 35 Van der Auwera, G.A., M.O. Carneiro, C. Hartl, R. Poplin, G. del Angel *et al.*, 2013 From fastQ  
36 data to high-confidence variant calls: the genome analysis toolkit best practices pipeline.  
37 *Current Protocols in Bioinformatics* 43 (1):11.10.11-11.10.33.
- 38 Waples, R.S., M.J. Ford, K. Nichols, M. Kardos, J. Myers *et al.*, 2022 Implications of large-  
39 effect loci for conservation: a review and case study with Pacific salmon. *Journal of*  
40 *Heredity* 113 (2):121-144.
- 41 Whiteman, J.P., H.J. Harlow, G.M. Durner, E.V. Regehr, B.C. Rourke *et al.*, 2017 Polar bears  
42 experience skeletal muscle atrophy in response to food deprivation and reduced activity  
43 in winter and summer. *Conservation Physiology* 5 (1).
- 44 Wickham, H., 2016 *ggplot2: elegant graphics for data analysis*: Springer-Verlag New York.
- 45 Wright, J.W., M.L. Stanton, and R. Scherson, 2006 Local adaptation to serpentine and non-  
46 serpentine soils in *Collinsia sparsiflora*. *Evolutionary Ecology Research* 8:1-21.

- 1 Yao, Y., and A. Ochoa, 2023 Limitations of principal components in quantitative genetic  
2 association models for human studies. *eLife* 12:e79238.
- 3 Yuan, F.L., A.H. Freedman, L. Chirio, M. LeBreton, and T.C. Bonebrake, 2018  
4 Ecophysiological variation across a forest-ecotone gradient produces divergent climate  
5 change vulnerability within species. *Ecography* 41 (10):1627-1637.
- 6 Zhou, X., and M. Stephens, 2012 Genome-wide efficient mixed-model analysis for association  
7 studies. *Nature Genetics* 44 (7):821-824.

## FIGURE CAPTIONS

**Figure 1.** The study area including the Central Coast and adjacent interior of British Columbia (BC). The coastal ecoregions are shown with a black hash, and grizzly bear sampling locations are indicated by orange points ( $n = 31$ , samples from 1996-2016). Grizzly Bear Population Unit (GBPU) boundaries are shown in teal, and the Bulkley and Tweedsmuir GBPUs are indicated with A or B, respectively. The yellow dashed boxes indicate the approximate areas of the Bella Coola Valley (South) and valleys containing Kitimat and Kemano (North). The inset situates the study area within North America.

**Figure 2.** (A) Samples clustered by principal components analysis (PCA) based on genotypes including all individuals ( $n = 31$ ) and linkage-filtered SNPs. Sampling site geographic locations are indicated by point size (latitude) and colour (longitude). (B) Relatedness distribution of pairs of all grizzly bears in the analysis. The hatched vertical line indicates Ritland metric of 0.15, the cutoff applied defining pairs with elevated relatedness.

**Figure 3.** ADMIXTURE fractions plotted per sample along with GPS coordinates. Pie charts are shown with their ADMIXTURE fractions, and those with a yellow ring outline were retained for outlier loci detection analysis. Open circles are samples that were removed due to high relatedness ( $n = 3$ ).

**Figure 4.** Manhattan plots for coastal vs. inland comparisons showing putative outlier SNPs using (A) DAPC, where red line indicates top 99.99<sup>th</sup> percentile; (B) pcadapt, where significance is Benjamini-Hochberg corrected  $p < 0.01$  based on PC2 only; and (C) GEMMA, where significance is Bonferroni MTC  $p \leq 0.05$ . Scaffolds with at least 100 SNPs present in full dataset are shown, with prefix (NW\_0266) removed. Black dots are highlighted SNPs that have been identified by multiple methods.

**Figure 5.** Top outlier region on genomic scaffold NW\_026622863, identified as having clusters of outlier loci as detected by all three methods near 47-48 Mbp, shown for (A) DAPC ( $n = 2$ ); (B) pcadapt (PC2 only;  $n = 96$ ); and (C) GEMMA ( $n = 11$ ). Significance cutoffs are described in the full chromosome Manhattan plot figure caption. Black dots are highlighted SNPs that have been identified by multiple methods.

**Figure 6.** Selected significant outlier SNPs consistently identified by multiple methods within the region of interest on scaffold NW\_026622863 shown as the number of alternate alleles for the genotype of each individual from the three identified regional groupings. NA values indicate missing genotype data for a sample.

**Table 1.** Genomic regions identified by multiple outlier detection methods with at least five loci on the same scaffold that were identified by at least two methods. Regions are shown with Mbp positions of the region and the number of shared outlier SNPs in the region in parenthesis (i.e., count). Genes within 10 kbp of the identified regions are shown, and those discussed in-text are underlined and acronyms given below the table. All significant outliers, shared outliers between methods, and gene acronyms are given in Additional File S2.

Scaffold	Methods	Shared loci	Positions Mbp (count)	Genes in region
NW_026622786	DAPC & pcadapt	11	18.56-18.59 (4); 25.96-25.98 (7)	DPYD (25.20-25.99)
NW_026622797	DAPC & pcadapt	5	14.79-14.93 (4); 25.49 (1)	CCDC170 (14.77-14.88), ARMT1 (14.90-14.91), RMND1 (14.91-14.95), U4 (14.93)
NW_026622863	DAPC & pcadapt	5	47.87-47.88 (2); 56.63 (3)	MARK4 (47.86-47.89), <u>CKM</u> (47.89-47.90); <u>OZF-like</u> (56.62-56.62), FRP2 (56.63-56.65)
NW_026622863	GEMMA & pcadapt	11	47.75-47.82 (11)	PPP1R37 (47.72-47.77), NKPD1 (47.77-47.77), TRAPPC6A (47.78-47.79), BLOC1S3 (47.79), EXOC3L2 (47.83-47.85)
NW_026622875	DAPC & pcadapt	83	19.86-19.97 (83)	<u>SLC9A9</u> (19.61-20.16)
NW_026622952	DAPC & pcadapt	7	28.50-28.51 (6); 28.60 (1)	WDR17
NW_026622997	DAPC & pcadapt	5	7.64-7.65 (5)	SERPINB1 (7.64-7.65)

CKM = *creatine kinase, M-type*; OZF-like = *zinc finger protein OZF-like*; SLC9A9 = *solute carrier family 9 member A9*.

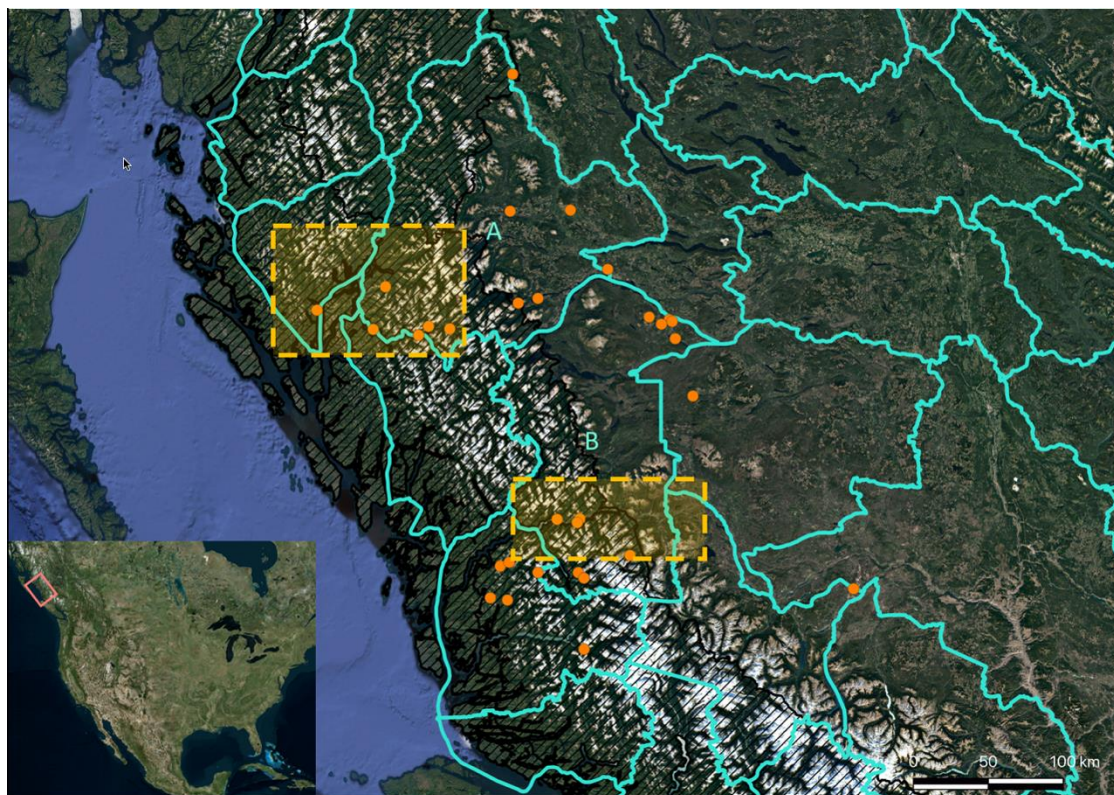
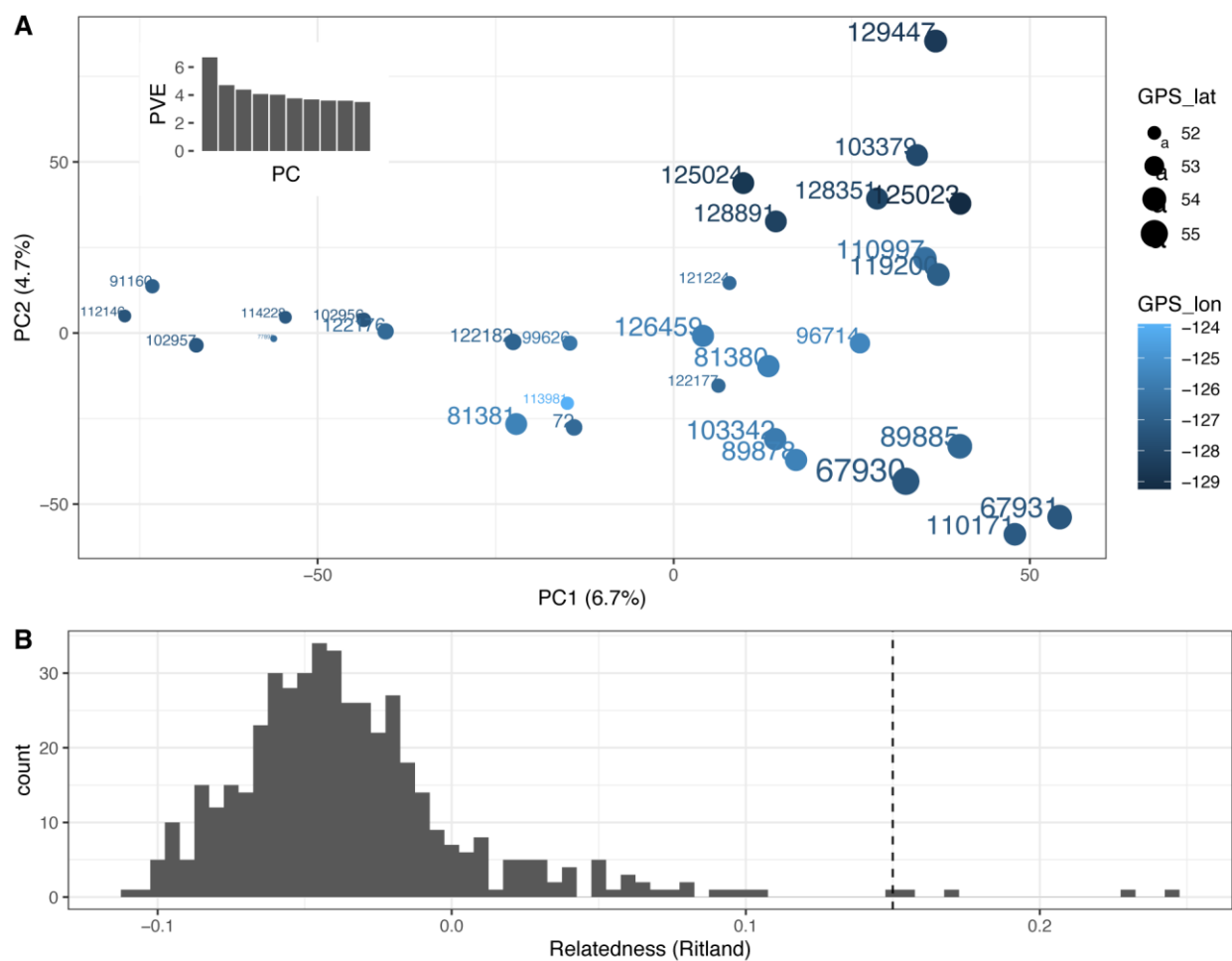
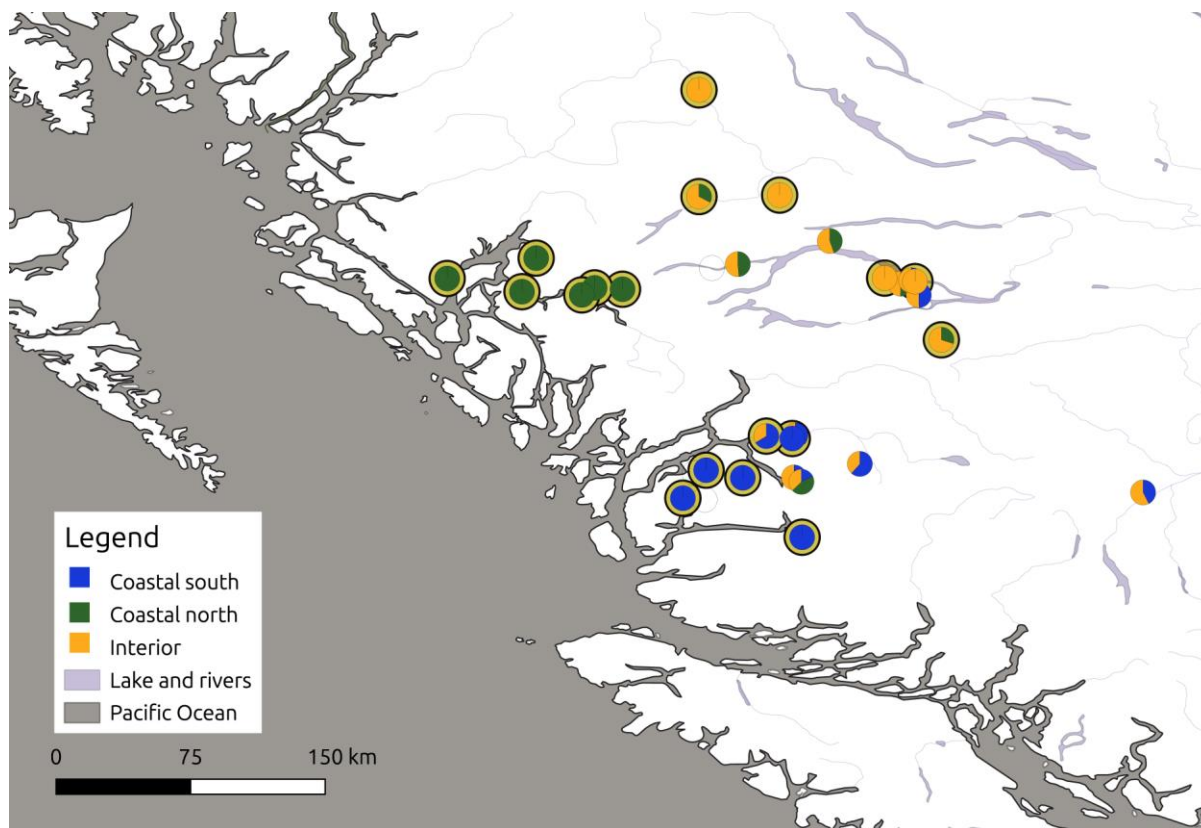


Figure 1  
148x105 mm ( x DPI)



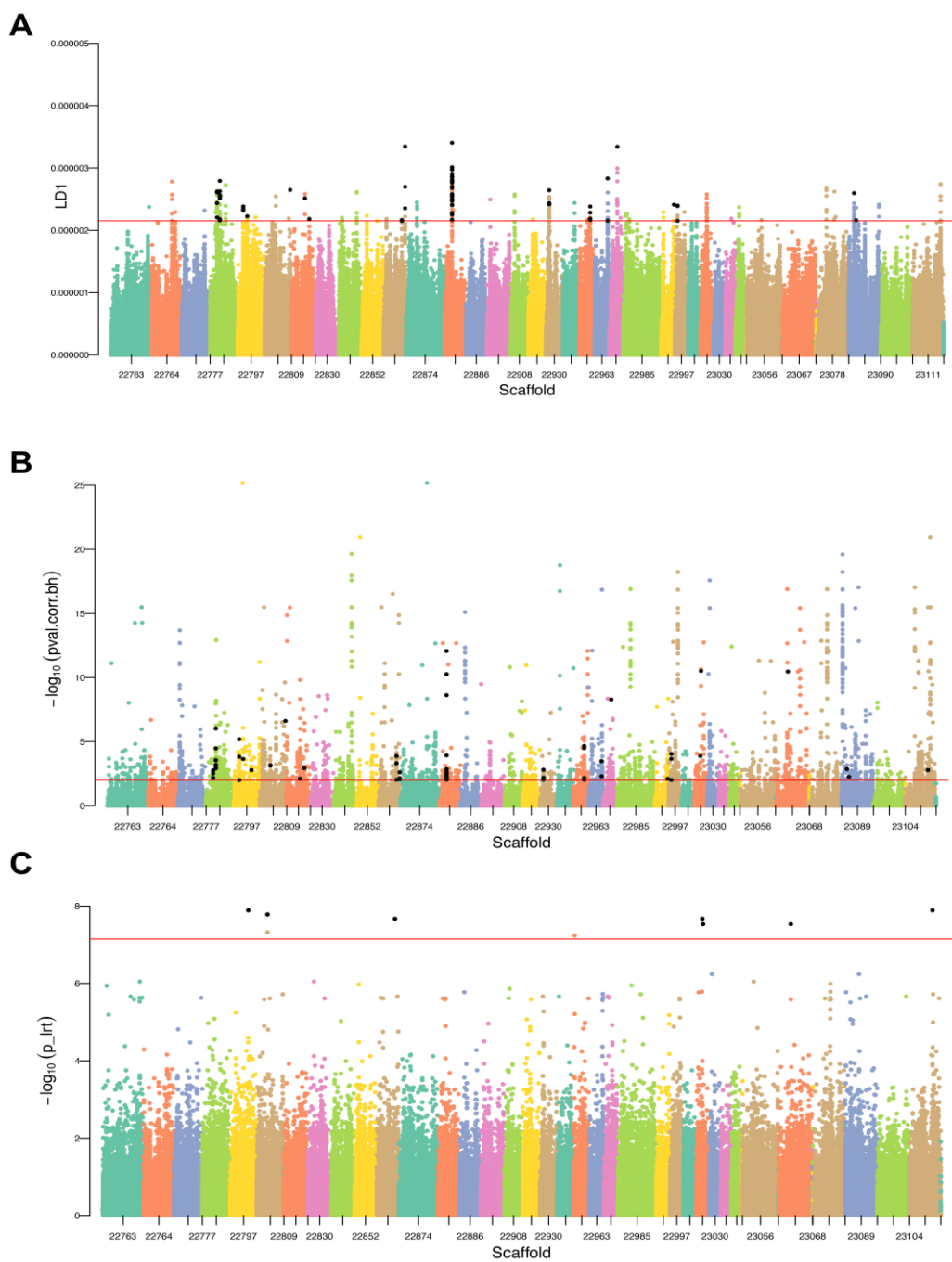
**Figure 2**  
165x128 mm ( x DPI)





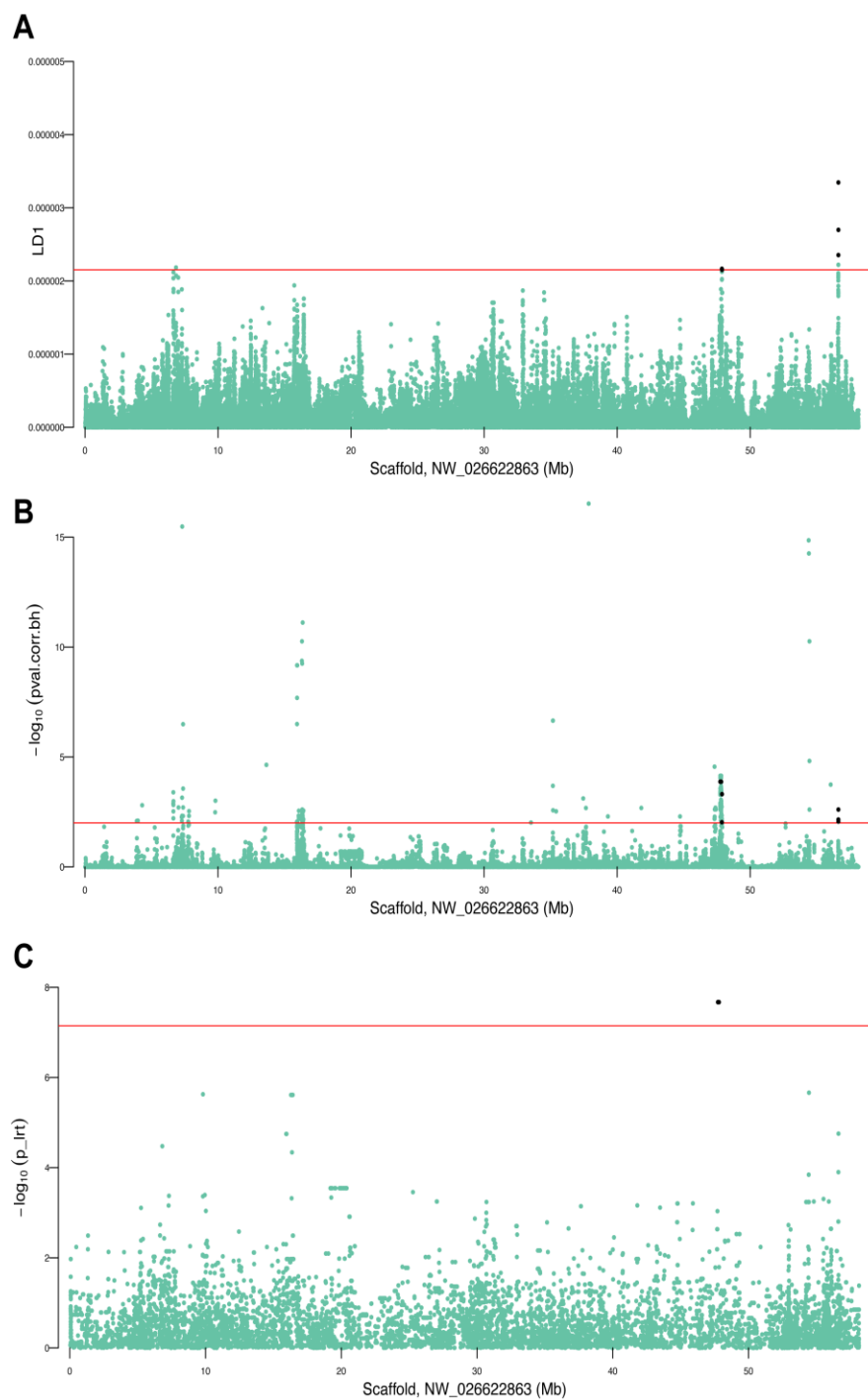
**Figure 3**  
159x108 mm ( x DPI)



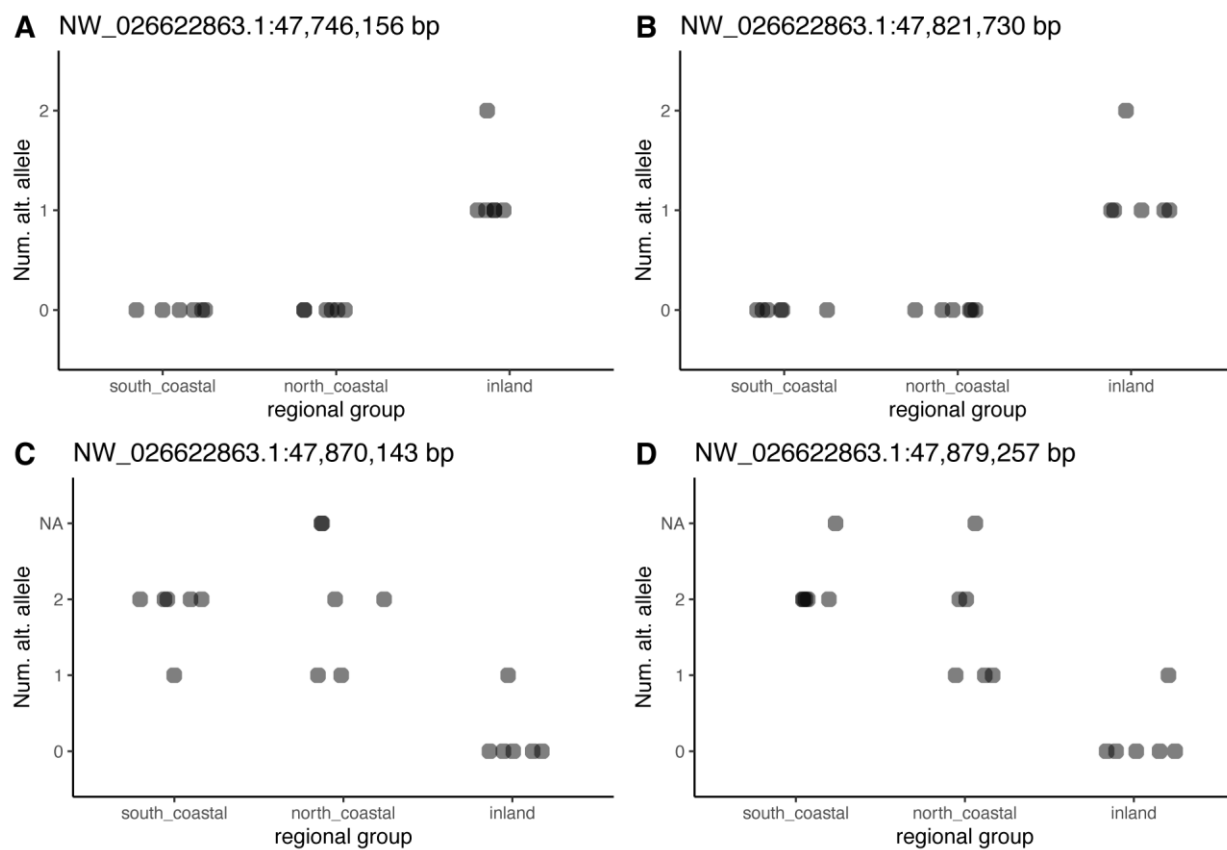


**Figure 4**  
236x341 mm ( x DPI)

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2  
3  
4



**Figure 5**  
236x341 mm ( x DPI)



**Figure 6**  
164x113 mm ( x DPI)